Summary of temporal trends in pollutant levels observed in marine mammals*

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ABSTRACT

The present paper reviews reported time trends in concentrations and relative abundance of pollutants in marine mammals. Available information refers only to pinnipeds and cetaceans, mainly covers the period 1969-1988 and focuses on DDTs, PCBs and mercury. Although data are limited, there are indications that in the Canadian Arctic, mercury levels in marine mammals have increased in recent decades. By contrast, during the late 1970s and the 1980s, concentrations of DDTs and PCBs in marine mammals from highly polluted areas have tended to decrease. While this trend is likely to continue for DDTs in the future, it is foreseen that until at least the first decades of the next century, PCB levels will stabilise as degradation is compensated by new inputs caused by the recycling of the fraction currently present in non-marine compartments.

KEYWORDS: REVIEW; TRENDS; POLLUTION-METALS; POLLUTION- ORGANO-CHLORINE; POLLUTION-PESTICIDES; CETACEANS-GENERAL; PINNIPEDS

INTRODUCTION

The history of the production, use and release into the environment of the different chemicals now considered as pollutants is complex and extremely variable from one compound to another. This fact, together with their different persistence and dispersal rates, makes it extremely difficult to assess historic time trends as well as to predict future trends in the level of exposure of cetaceans to these compounds. For synthetic chemicals such as organochlorines and organobrominates, it is obvious that present concentrations are higher than in preindustrial times, although variations over time in recent years are not easy to determine. In the case of trace elements, radionuclides and polyaromatic hydrocarbons, compounds that are also naturally occurring, there is some controversy as to whether the levels detected in the environment today have always been there. The general opinion is that pre-industrial levels were lower and that an increasing temporal trend is superimposed on the geochemical baseline levels (Wagemann *et al.*, 1990).

It is the intention of this summary paper to delineate observed trends in pollutant levels in marine mammals to assess probable future trends on a global scale.

HISTORIC TRENDS

For synthetic chemicals, there is a paucity of data on levels in pre-industrial or early industrial times. This is largely due to the fact that most compounds of this type presently found in marine mammals were only introduced in the 1930s or later.

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In the case of trace elements, the situation is different. Glacial records show that heavy metals have been in the marine environment since prehistoric times (Murozumi *et al.*, 1969; Weiss *et al.*, 1971). The natural cycling of mercury for example, has usually exceeded that of anthropogenic origin. For this reason, in many cases, trace element tissue concentrations in wildlife can be assumed to be natural, with the exception of certain localised cases, mainly in enclosed seas or riverine systems However, diverse sources, ranging from geological registers and the composition of polar ice to historic autopsy studies, suggest that the levels of exposure to at least some heavy metals may have increased markedly in modern times This holds particularly for concentrations of mercury, lead and cadmium, which have increased since the beginning of the present century in both marine mammals and man (Wagemann *et al.*, 1990). A noteable example is the increase in the lead content of Greenland ice, which is today 200-times that considered 'natural' some thousand years ago (Goyer, 1991).

TRENDS SINCE 1965 AND PROSPECTS

Table 1 summarises our literature survey of pollutant trends observed in marine mammals (usually refereed publications have been considered and the results as presented by the authors have been considered to be valid). About 60% of the studies refer to Pinnipedia and the rest to Cetacea. The period covered ranges from 1965 to 1994, although most data correspond to the period 1969-1988. The apparent decrease in information after 1988 may reflect the time-lag for publication of more recent surveys. Almost all information is for the Northern Hemisphere. Only two surveys (bottlenose dolphins and common dolphins in South Africa) were carried out during the period 1980-1987 in the Southern Hemisphere.

Data on pollutants other than PCBs and tDDT are insufficient to evaluate trends on a global scale. However, an interesting example of a local trend is the increase in liver mercury levels reported for white whales in the western and eastern Canadian Arctic between the early 1980s and 1994, in ringed seals in the western Canadian Arctic and in narwhals in the eastern Canadian Arctic (Wagemann *et al.*, 1996). Apparently, there was no temporal change in cadmium levels in these species during this period.

Results for PCBs and tDDT for the period prior to 1977 are highly variable and include increases in some areas, although in most cases no significant trends were observed.

Since 1977, all studies have reported either no trend or a decrease in tissue pollutant levels. This may point to a continuing reduction of exposure to contaminants in most marine mammal populations. In relative terms, a decrease in concentrations was first noticed for DDT and later for PCBs, a fact which is consistent with the history of the production, use and restriction of these two groups of organochlorine compounds and their persistence in the ecosystem (de Voogt and Brinkman, 1989; Peterle, 1991). The decrease in DDT levels caused by the discontinuation in its use in pesticides has been associated in many areas with a parallel increase in the relative abundance of its metabolised forms. This has led to the use of various ratios between metabolised and parental compounds of DDT, to assess the time passed since the last inputs of the pesticide were introduced into the ecosystem. In particular, during the 1970s and 1980s, the ratio DDE/tDDT was found to progressively rise in odontocetes and pinnipeds in the North Atlantic (Aguilar, 1984) and in grey and harp seals from eastern Canada (Addison et al., 1984). The decrease of PCBs was particularly apparent in areas close to sources such as Lake Ontario, the Baltic Sea, the Wadden Sea and the North Sea (OECD, 1980; Addison et al., 1986; Olsson and Reutergård, 1986; Reijnders, 1996a). However, concentrations in marine biota from the Baltic and the North Seas levelled off in the 1980s (de Boer, 1988; 1994; Bignert et al., 1993) and similar observations were made by Norstrom et al. (1988) for polar bears in the Canadian Arctic, and by Tanabe et al. (1994b)

Area	Species	Period	PCB	tDDT	HCB	HCHs	Diel	Hg	Reference
BALTIC									
Baltic Sea	H. grypus	1969-88	nt	d					Blomkvist et al.,1992
									Olsson et al.,1994
Aland Sea	H. grypus	1970,	nt	nt					Olsson et al.,1974
		1971-72,							
	B 1 1 1 1	1973							
Baltic Sea	P. hispida	1969,	d	d					Blomkvist et al., 1992
		1973-80,							Olsson et al.,1994
Gulf of Bothnia	P. hispida	1988		-					01
Gun or Bouina	r . mspiuu	1969, 1971-72,	nt	nt					Olsson et al.,1974
		1971-72,							
Gulf of Bothnia	P. hispida	1972-77	i	nt					Helle,1981
Gulf of Bothnia	P. hispida	1977-80	d	d					Helle,1981
Gulf of Bothnia	P. hispida	1980-83	nt	nt					Helle and Stenman,
	•								1984
Lake Saimaa	P. hispida	1970-1977,	d	d					Helle et al,1983
		1981							,
Gulf of Finland	P. hispida	1977,	d	d					Helle et al.,1985
		1980-83							
NORTH SEA									
Farne Islands	H. grypus	1968-75	nt	d					Holden,1978
Scottish coast	H. grypus	1965-71	nt	d			d		Holden,1975
Dutch coast	P. vitulina	1973-81	d						Van der Zande and de
Dutch coast	P. vitulina	1969,						nt	Ruiter,1983 Koeman <i>et al.</i> ,1972;
	1	1970-75,						m	Reijnders,1980
		1976							Reijhuers,1960
Dutch coast	P. vitulina	1973,	d						Reijnders, 1980;
		1975-88							Reijnders, 1996a
E. Canada									5
Sable Island	H. grypus	1974,	nt	d					Addison et al.,1984
		1976-82							
North Baffin	P. hispida	1972,	d	d		nt	nt		Muir et al.,1988
		1976-84							
	P. groenlandicus		d	d			nt		Jones et al., 1976
	P. groenlandicus		d	d			i		Ronald et al., 1984
	P. groenlandicus		nt	d					Addison et al.,1984
St Lawrence Gulf NORTHEAST USA	P. groenlandicus	1982-89	d	nt					Beck et al., 1993
Northeast USA	P. vitulina	1980-90,	d	d	nt			nt	Lake <i>et al.</i> ,1995
New TOIK	r. vitutina	1980-90,		(DDE)	ш			ш	Lake et ut., 1995
Canadian Arcti	с	. / / 6							
Holman Island	P. hispida	1972-81	d	d					Addison et al.,1986
Western Arctic	P. hispida	1987-93		-					Wagemann et al., 1996
Eastern Arctic	P. hispida	1989-94							Wagemann et al., 1996
N. NORTH PACIFIC	•								g
Japan	C. ursinus	1971-76	i	i		i			Tanabe et al.,1994b
Japan	C. ursinus	1976-88	d	d		d			Tanabe et al., 1994b

 Table 1(a)

 Temporal trends in pollutant levels in pinnipeds, i: increase, d: decrease, nt: no trend.

Area	Species	Period	PCB	tDDT	HCB	HCHs	Diel	Hg	Reference
E. CANADA									
W. Hudson Bay	D. leucas	1966, 1967-86	nt	d	nt	i			Muir et al. (1990)
Bay of Fundy	P. phocoena	1969-73		d					Gaskin et al. (1982)
Bay of Fundy	P. phocoena	1971-77	nt						Gaskin et al. (1983)
Bay of Fundy	P. phocoena	1969-73						d	Gaskin et al. (1979)
Bay of Fundy	P. phocoena	1974-77						i	Gaskin et al. (1979)
CANADIAN ARC									
Western Arctic	D. leucas	1981, 1984-93, 1994						i	Wagemann et al. (1996)
Eastern Arctic	D. leucas	1984-93, 1994						i	Wagemann et al. (1996)
Western Arctic	M. monoceros	1978, 1979- 92, 1994						i	Wagemann <i>et al.</i> (1996)
N. NORTH PACIF	FIC								
Japan	S. coeruleoalba	1978, 1979-86	nt	nt	d	d			Loganathan <i>et al.</i> (1990)
SOUTHERN AFRI	CA								
East coast	T. truncatus	1980, 1983-84, 1987	nt	nt			nt		Cockroft <i>et al.</i> (1989)
East coast	T. truncatus	1980-87	nt	d					Kock et al. (1994)
East coast	D. delphis	1980-85	nt	nt					Kock et al. (1994)
Mediterranea	•								
N.W. coast NORTH SEA	S. coeruleoalba	1987-94	d	d					Borrell et al. (1996)
Scottish coast	P. phocoena	1965-71	nt	nt			d		Holden (1975)
	B. acutorostrata	1984-91	i	nt	nt	nt			Tanabe et al. (1995)

 Table 1(b)

 Temporal trends in pollutant levels in cetaceans, is increase, di decrease, nt: no trend

for northern fur seals in the Pacific. Apart from the apparent levelling off of the decrease of PCBs in local populations, Tanabe (1988) concluded that the global PCB levels are unlikely to decline in the near future due to the fact that only 30% of all the PCBs produced have been dispersed into the environment. In this context, we can also consider the estimation of Bletchly (1984), that disposal of PCBs will peak at the end of the 1990s.

With respect to future trends in levels in marine mammals, Tateya *et al.* (1988) predicted, based on studies in striped dolphins, that levels of PCBs in marine mammals would be at their highest between 2000 and 2030.

Given the fact that of the *ca* 2,000,000 tonnes of PCBs produced, only 1% has reached the ocean (Reijnders, 1996b), slow dispersal will continue and it is expected that on a global scale, no apparent reduction in the exposure of marine mammals to PCBs will occur until the turn of the 21st century.

In view of the presumed global change in the distribution of organochlorine residue levels, the Arctic waters and adjacent seas and oceans are expected to become the major sink for these contaminants (Tatsukawa, 1993). This holds to a much lesser extent for the Southern Hemisphere (Tanabe *et al.*, 1994a). It is important that monitoring programmes for pollutant trends in marine mammals take this into account.

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